

# Simulation of nitrate aerosol concentrations over East Asia with the model system RAMS-CMAQ

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(Manuscript received 25 April 2006; in final form 1 November 2006)

## ABSTRACT

The Models-3 Community Multi-scale Air Quality modeling system (CMAQ) coupled with the Regional Atmospheric Modeling System (RAMS) is used to analyse the geographical and the seasonal characteristics of nitrate aerosol ( $\text{ANO}_3$ ) concentration distributions over East Asia. Three-dimensional concentrations in January, March, April, July and October of 2001 are simulated, and for the evaluation of model performances, the simulated values of wind direction, wind speed, temperature, specific humidity, nitric oxide (NO), nitrogen oxide ( $\text{NO}_2$ ),  $\text{ANO}_3$  and ammonium aerosol ( $\text{ANH}_4$ ) are compared with the observational data obtained onboard the P-3B aircraft during its four consecutive flights over the South China Sea, the East China Sea, the Yellow Sea and the water areas to the east of Japan conducted on 13, 17, 18 and 21 March 2001. The observed data are 5-min averaged and the model results with a 1-hour temporal resolution are interpolated to the aircraft location and time using trilinear interpolation. Comparison shows that the observed values exhibit strong temporo-spatial variations, and the model reproduces these variations reasonably well. The simulated values of wind direction, wind speed, temperature and specific humidity are generally in good agreement with the observed ones, their correlation coefficients reach 0.66, 0.94, 0.99 and 0.96, respectively. Comparison also shows that the average mixing ratios of modelled and observed NO,  $\text{NO}_2$ ,  $\text{ANO}_3$  and  $\text{ANH}_4$  agree reasonably well with each other, the correlation coefficient for  $\text{ANO}_3$  and  $\text{ANH}_4$  are larger than 0.8, but the simulated standard deviations are smaller than the observed one, and the correlation coefficient for NO and  $\text{NO}_2$  are 0.48 and 0.44.

Analysis of horizontal distributions of monthly averaged  $\text{ANO}_3$  concentrations in the boundary layer indicates that the  $\text{ANO}_3$  mixing ratios have noticeable differences among the four seasons. Generally the concentrations are high in the winter, spring and fall and low in the summer, and seasonal variations are typically strongest over China and Japan. Highest concentrations are found over eastern China where emissions of nitrogen oxides and ammonia are high and long-range transport may cause elevated concentrations in remote areas under favourable meteorological conditions, but the seasonal variation in the  $\text{ANO}_3$  concentrations is heavily influenced by the changes in precipitation and temperature.

## 1. Introduction

Nitrate aerosols ( $\text{ANO}_3$ ) are secondary aerosols produced from nitrogen oxides ( $\text{NO}_x = \text{NO} + \text{NO}_2$ ), and play an important role in many environmental issues as climate changes, acid rain and smog. A recent critical assessment of nitrate observations in Europe showed that nitrate significantly contributed to the aerosol concentration in Northern Europe, especially in winter (October–March), large contributions of nitrate to the total aerosol mass were found in the western Europe, where nitrate concentrations often exceeded those of sulphate ( $\text{ASO}_4$ ) (Schaap

et al., 2002). In the United Kingdom  $\text{ANO}_3$  concentrations were reported to have a southeast to north-west gradient, and its annual average levels in the south-east region were generally higher than those of  $\text{ASO}_4$  (AQEG, 2005; Turnbull and Harrison, 2000). While in East Asia, a region of the world with large and rapidly increasing anthropogenic emissions, the  $\text{NO}_x$  emissions have increased by 58% from 1975 (2.05 TgN/yr) to 1987 (3.25 TgN/yr) (Kato and Akimoto, 1992), and van Aardenne et al. (1999) predicted an increase of almost fourfold in  $\text{NO}_x$  emissions from 1999 to 2020. Especially in China anthropogenic emissions associated with fossil fuel burning have grown significantly as there experienced a period of rapid economic development and industrial expansion in the last three decades (e.g., Streets and Waldhoff, 2000). Many observations indicate that the increase in  $\text{NO}_x$  emissions is exacerbating the acid rain problem in East

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DOI: 10.1111/j.1600-0889.2007.00255.x

Asia (e.g., Kim and Cho, 2003), especially in China (e.g., Xie, 2002; Yu et al., 2003). Xie (2002) reported that nitrate ion was steadily increasing in the rain water collected in Guangzhou in southern China in the period of 1996 to 2000.

At present our knowledge of aerosol distributions and associated effects on air quality and climate changes is largely dependent on models (IPCC, 2001). There are few observational data from remote locations against which such models can be tested, especially over the oceans (IPCC, 2001). Such data are especially critical in areas impacted by transport from regions with rapidly developing economies where emission controls are often minimal and emission rates poorly documented. So although nitrate aerosols were often included in the calculations of regional-scale chemical transport models applied to East Asia (e.g., Zhang et al., 2004; Han et al., 2005), their temporo-spatial distributions are still poorly characterized over the region.

In the period from 4 March to 2 April 2001, the Transport and Chemical Evolution over the Pacific (TRACE-P) field campaign (Jacob et al., 2003) was conducted with two NASA aircrafts (DC-8 and P-3B) over the western Pacific along the Asian Pacific Rim. The aircrafts DC-8 and P-3B performed 11 and 12 observations, respectively, with bases near Hong Kong, Okinawa and Tokyo, and each flight lasted about 9 hours. These flights are named as DC-8 07 to 17 and P-3B 08 to 19. These airborne observations obtained an extensive suite of meteorological parameters, trace gases and aerosols. This study is to take advantage of this observational data set to examine regional-scale performances of a comprehensive chemistry-transport model system (Zhang et al., 2004) and then to investigate geographical and seasonal characteristics of nitrate aerosol concentration distributions over East Asia.

This paper is divided into four sections. We briefly describe the model, its initial and boundary conditions in section 2. In section 3, we firstly compare model results with the observations from the TRACE-P missions, and then discuss geophysical and seasonal distributions of nitrate aerosol concentrations over East Asia. Conclusions are presented in section 4.

## 2. Model system and input parameters

The model system used in this study has two major components - the Models-3 Community Multi-scale Air Quality (CMAQ) modeling system (Byun and Ching, 1999) and the Regional Atmospheric Modeling System (RAMS; Pielke et al., 1992). CMAQ is an Eulerian-type model and it simulates concurrently the atmospheric and land processes affecting the transport, transformation and deposition of air pollutants and their precursors on both regional and urban scales. While RAMS is a highly versatile numerical code for simulating and forecasting meteorological phenomena. In this study it is used to produce regional scale meteorological fields including boundary layer turbulence, cloud and precipitation necessary for the CMAQ calculation. CMAQ coupled with RAMS has recently been successfully applied to

East Asia to simulate tropospheric ozone and sulphur dioxide (Zhang et al., 2003, 2004)

In this study CMAQ is configured with the chemical mechanism of the Regional Acid Deposition Model version 2 (RADM2; Stockwell et al., 1990), including gas-phase and aqueous reactions leading to secondary aerosol formation, and extended to include aerosol processes on the basis of the Regional Particulate Model (RPM; Binkowski and Shankar, 1995). In the extended RADM2 the particle size distribution is represented as the superposition of three lognormal subdistributions, and the processes of coagulation and particle growth by the addition of new mass, particle formation, etc. The reaction of nitrogen pentoxide ( $\text{N}_2\text{O}_5$ ) on aerosol surface is a source for nitric acid during night-time, whereas during the day the nitrogen trioxide radical is readily photolysed. This reaction is parameterized following Dentener and Crutzen (1993). The thermodynamic equilibrium between gaseous nitric acid ( $\text{HNO}_3$ ), ammonia ( $\text{NH}_3$ ) and particulate ammonium nitrate ( $\text{NH}_4\text{NO}_3$ ) and ammonium sulphate and aerosol water is considered (Kim et al., 1993); the ammonia deficient regime (in which the ionic ratio of ammonium to total sulphate ion is less than two) leads to an acidic aerosol system with very low concentrations of dissolved nitrate ion which depend very strongly on ambient relative humidity; the second regime is one in which the ammonium to sulphate ratio exceeds two, the sulphate is completely neutralized, and there is excess ammonia. If there is nitric acid vapour in the system, it will dissolve in the aqueous particles along with the excess ammonia and produce abundant nitrate. For the cases when the relative humidity is low that the aerosol liquid water content comprises less than 20 percent of the total aerosol mass, and the ionic ratio of ammonium to sulphate is greater than two, the appearance of "dry ammonium nitrate" aerosol is calculated from the following equilibrium relationship (Mozurkewich, 1993):



For the CMAQ inputs, emissions of nitrogen oxides, sulphur dioxide, carbon monoxide, non-methane hydrocarbon species and ammonia from anthropogenic activities and open biomass burning were obtained from the  $0.5^\circ \times 0.5^\circ$  monthly inventory (Streets et al., 2003). Nitrogen oxides and ammonia from soil were adopted from the Global Emissions Inventory Activity (GEIA)  $1^\circ \times 1^\circ$  monthly global inventory (Benkovitz et al., 1996). Aircraft emissions were based on the Emission Database for Global Atmospheric Research (EDGAR; Olivier et al., 1996). Sulphur dioxide emissions arising from volcanoes were based on the estimates by Streets et al. (2003). These different emission inventories are preprocessed to form a unified one with the same spatial resolution of CMAQ for the simulation.

In this study RAMS was exercised in a four-dimensional data assimilation mode using analysis nudging with re-initialization every four days, leaving the first 24 hours as the initialization period. The three-dimensional meteorological fields for RAMS were obtained from the European Center for Medium-Range

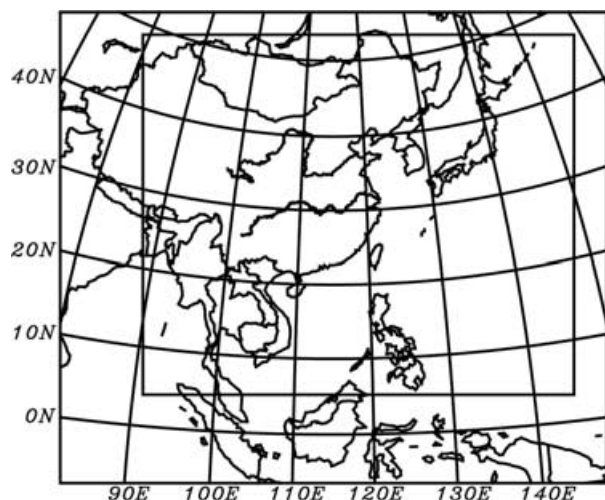


Fig. 1. Model domain for RAMS (outer region) and CMAQ (inside region) used in this study.

Weather Forecasts (ECMWF) analyzed data sets with  $1^\circ \times 1^\circ$  resolution and with six-hour interval. Sea Surface Temperatures (SST) for RAMS were based on weekly mean values and observed monthly snow-cover information as the boundary conditions for the RAMS calculation.

The model domain (shown in Fig. 1) is  $8000 \times 5600 \text{ km}^2$  (outside region) for RAMS and  $6240 \times 5440 \text{ km}^2$  (inside region) for CMAQ on a rotated polar-stereographic map projection centred at  $(25^\circ\text{N}, 115^\circ\text{E})$  with 80 km mesh. RAMS and CMAQ have the same model height. For RAMS there are 23 vertical layers in the  $\sigma_z$  coordinates system unequally spaced from the ground to  $\sim 23 \text{ km}$ , with about 9 layers concentrated in the lowest 2 km of the atmosphere in order to resolve the planetary boundary layer, while there are 14 levels for CMAQ with the lowest 7 layers being the same as those in RAMS.

Initial and boundary conditions of chemical species in CMAQ were chosen to reflect the East Asian situation. Recent measurements were used whenever possible. To evaluate the impact of the anthropogenic emissions on the distributions of trace gases and aerosols, the initial and boundary conditions were generally chosen at the lower end of their observed range (e.g., the northern and western boundary conditions for  $\text{O}_3$ ,  $\text{NO}_2$ ,  $\text{ANO}_3$  and  $\text{ASO}_4$  were 30 ppbv, 0.2 ppbv,  $0.5 \mu\text{g}/\text{m}^3$  and  $1 \mu\text{g}/\text{m}^3$ , respectively) so as to allow the emissions and chemical reactions to bring them closer to their actual values during the initialization period (Zhang et al., 2003).

### 3. Results and discussion

#### 3.1. Model evaluation

With the model system described in the previous section, meteorological fields and three-dimensional concentrations of trace gases and aerosols over East Asia in the periods of 26 December

2000 to 31 January, 22 February to 5 May, 25 June to 31 July and 25 September to 31 October of 2001 are simulated. Zhang et al. (2006) compared simulated concentrations of ozone and its related species with observations obtained during the TRACE-P experiment and comparison indicates that the model system reproduced the tempo-spatial distributions of ozone and related species reasonably well and most model results were within a factor of two of the observations. During the period of 4 March to 2 April, two aircrafts DC-8 and P-3B conducted 11 (DC-8 flights 07 to 17) and 12 (P-3B flights 08 to 19) observations, respectively, over a broad area covering Hong Kong, Okinawa, the East China Sea, the Yellow Sea and the water areas to the south and the east of Japan with bases near Hong Kong, Okinawa and Tokyo, and each flight lasted more than 8 hours (Jacob et al., 2003). In this study, four consecutive P-3B flights 12 to 15 were selected to evaluate the model performances. Their horizontal flight tracks are presented in Fig. 2. These flights were performed on 13, 17, 18 and 21 March, respectively, over four different areas: the South China Sea, the East China Sea, the Yellow Sea and the water areas to the east of Japan. Comparison with these specific flights is to test the model's ability to produce three-dimensional distributions of meteorological parameters and chemical species in more detail. For these comparisons the 5-min merged data set for the P-3B was used, and the model results with a 1-hour temporal resolution were interpolated to the aircraft location and time using trilinear interpolation.

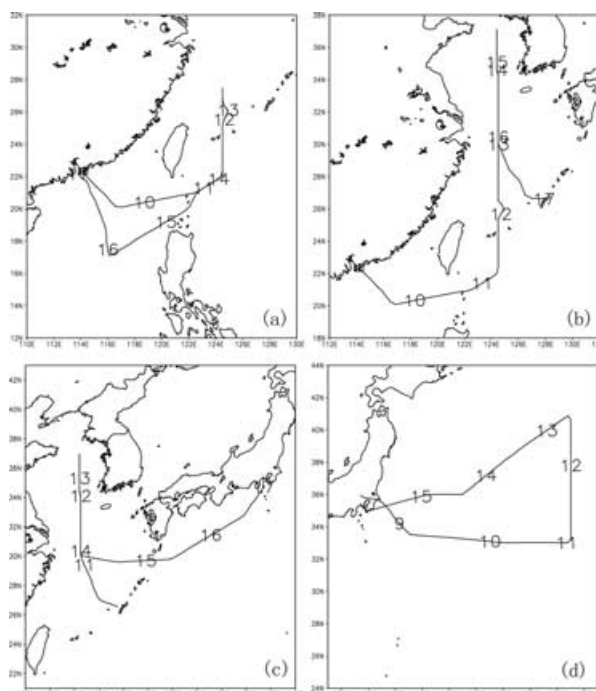


Fig. 2. Horizontal tracks of the P-3B flights 12 to 15 conducted on (a) 13 March, (b) 17 March, (c) 18 March, and (d) 21 March, respectively. Numbers are flight time in the Chinese Standard Time (CST).

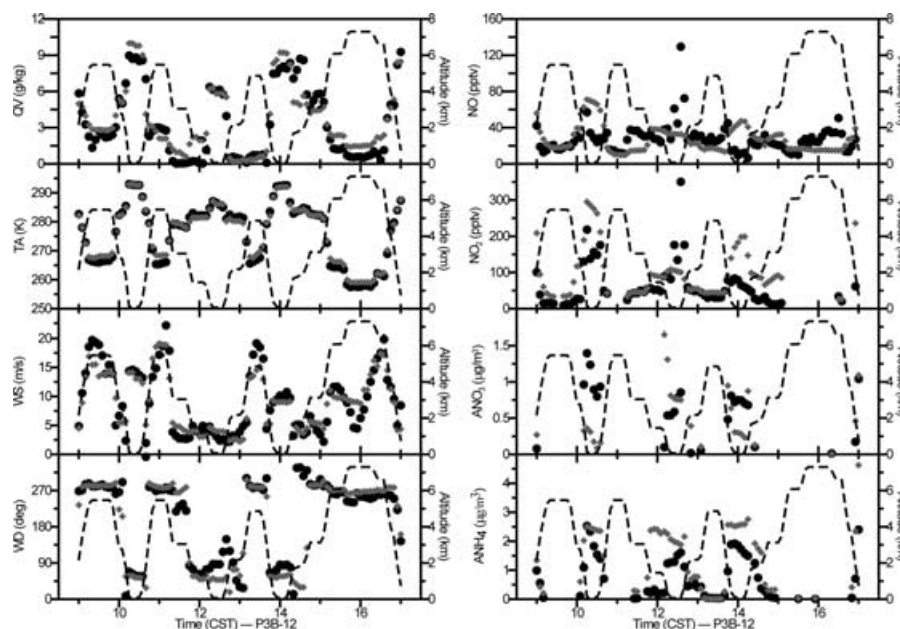


Fig. 3. Time series of observed (black dots) and simulated (grey diamonds) values of (a) specific humidity, (b) temperature, (c) wind speed, (d) wind direction, (e) NO, (f) NO<sub>2</sub>, (g) ANO<sub>3</sub>, and (h) ANH<sub>4</sub> along the track of the P3B flight 12 (dashed line, km) conducted on 13 March 2001. Units are g/kg, K, m/s, degree, pptv, pptv,  $\mu\text{g}/\text{m}^3$ , and  $\mu\text{g}/\text{m}^3$ , respectively.

Figures 3–6 show time series of observed and simulated values of specific humidity (QV), temperature (TA), wind speed (WS) and wind direction (WD) and concentrations of NO, NO<sub>2</sub>, ANO<sub>3</sub> and ammonia aerosol (ANH<sub>4</sub>) along the tracks of P-3B flight 12–15. On 13 March, the P-3B took off at about 0900 CST (Chinese Standard Time), i.e., 0100 UTC, from Hong Kong, and firstly headed south eastward and eastward, and then took a

track heading north to the east of Taiwan along 124.5°E. After about 1230 CST it flew back to Hong Kong with a short leg of high altitude sampling into the South China Sea planned to look for potential biomass outflow (*cf.* Figs. 2a and 3e–h). The observed meteorological parameters displayed in Figs. 3a–d show that specific humidity and air temperature decreased with height while wind speed increased, and westerly winds prevailed in the

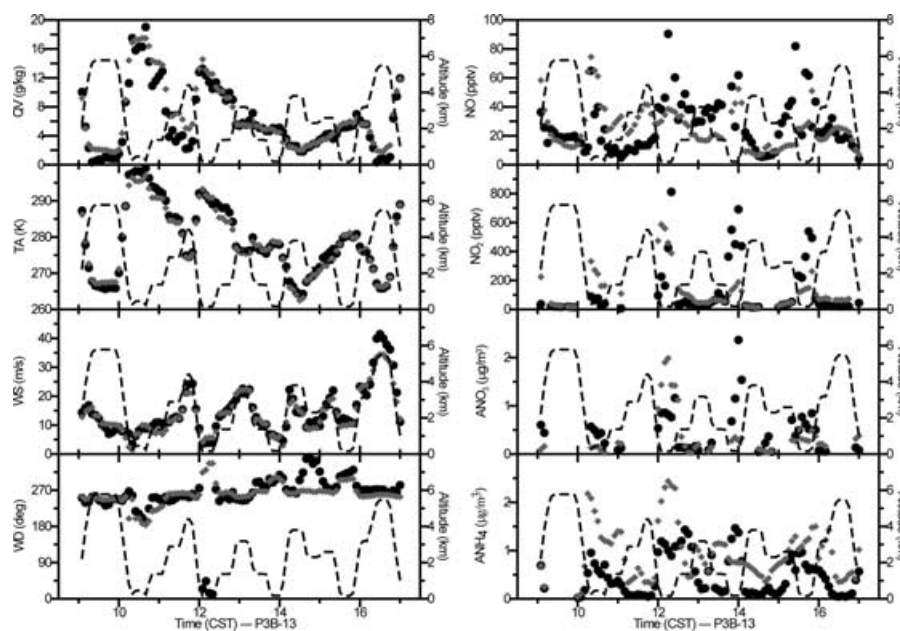


Fig. 4. Same as in Fig 3 but for the P3B flight 13 on 17 March 2001.

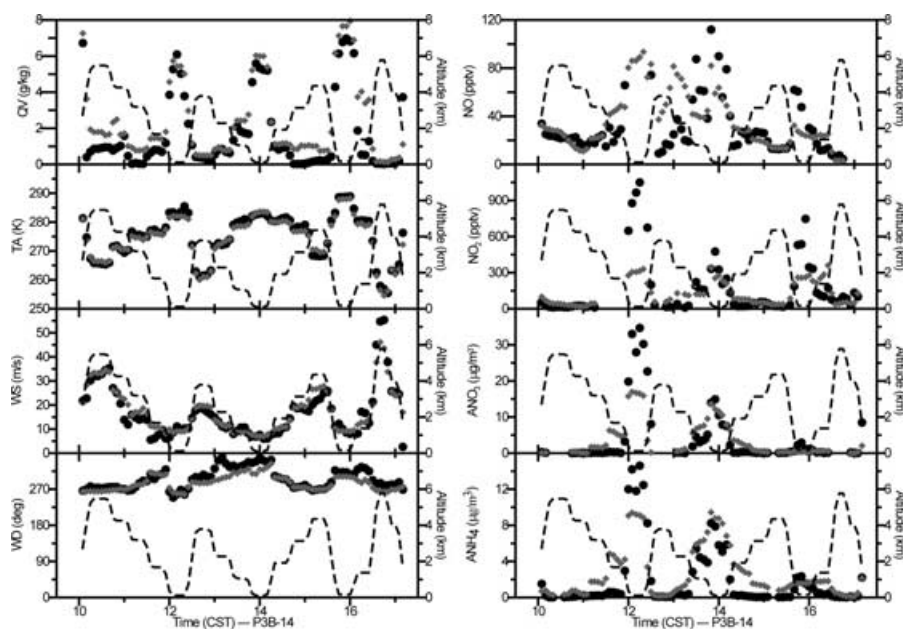


Fig. 5. Same as in Fig 3 but for the P3B flight 14 on 18 March 2001.

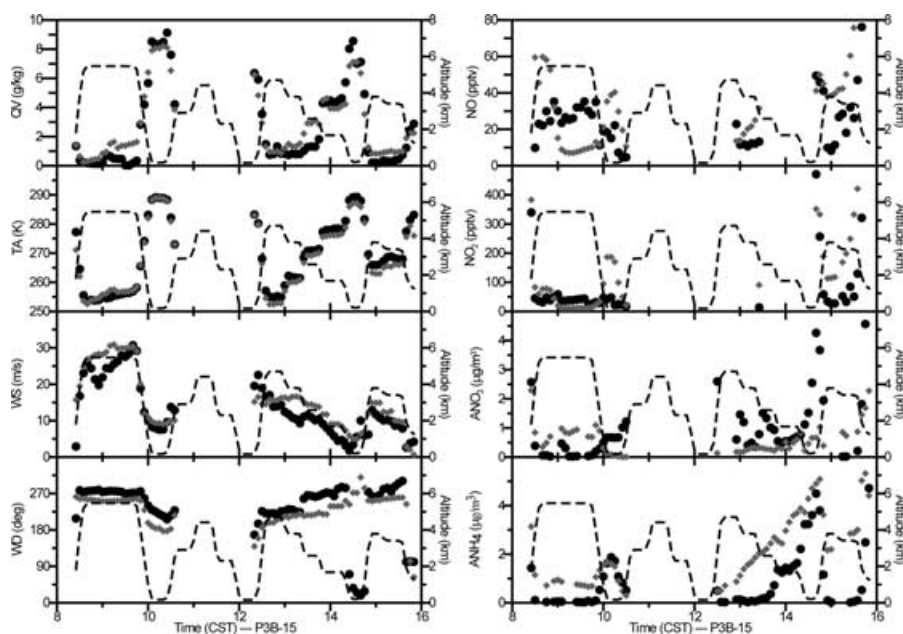


Fig. 6. Same as in Fig 3 but for the P3B flight 15 on 21 March 2001.

layers above  $\sim 4$  km while wind directions in the lower atmosphere changed from north-easterly in the area to the south of Taiwan ( $\sim 1030$  CST) to south-easterly over the East China Sea ( $\sim 1230$  CST). From Figs. 3a–d we find that the modelled values of specific humidity, temperature, wind speed and wind direction are generally in good agreement with the observed ones.

Figures 3e–h show that the measured  $\text{NO}$ ,  $\text{NO}_2$ ,  $\text{ANO}_3$  and  $\text{ANH}_4$  concentrations exhibited large variations in time and space: for example, their values were generally high in the lower

troposphere and usually low at upper layers, but differed sometimes greatly even at the same height. From these figures we find that the modelled values agree reasonably well with their observed ones, and the model captured their variations in time and space well, e.g., high concentrations of  $\text{NO}$ ,  $\text{NO}_2$  and  $\text{ANH}_4$  over the water area to the south of Taiwan ( $\sim 1030$  CST) were well reproduced. Preliminary analysis of model results indicates that these high concentrations might be associated with pollution outflow from Taiwan as north-easterly winds were simulated

over the area (*cf.* Fig. 3d). Figures 3e–f also show that the model could not capture very high levels of NO and NO<sub>2</sub>. Underestimation of maximum NO and NO<sub>2</sub> values may be primarily due to the coarse grid resolution and uncertainties in the emissions used in this study.

The P-3B started observations at about 0900 CST from Hong Kong on 17 March. It firstly flew south-eastward and then north-eastward to (22°N, 124.5°E), turned north at this point into the Yellow Sea along 124.5°E up to 37°N, and finally returned south into Okinawa. Figures 4a–d show that specific humidity and temperature in the lower atmosphere were higher at low latitudes (~1030 CST) than at mid-latitudes (~1400 and 1540 CST), winds were predominantly westerly in the layers above ~4 km and south-easterly in the low atmosphere at low latitude (~1030 CST) while north-westerly at mid-latitude (~1540 CST), and wind speeds at high altitudes were larger at mid-latitudes than at low latitudes. As shown in these figures, these observed features are all well reproduced by the model.

Figure 4d shows that prevailing winds were north-westerly in the lower atmosphere and at mid-latitude, so mixing ratios of many chemical species over there were under direct influences of the emissions over the continental boundary layer. From Figs. 4e–f we find quite high values of NO and NO<sub>2</sub> in the lower troposphere at mid-latitudes and their concentrations were generally higher than those at low latitudes monitored during the previous flight (*cf.* Figs. 3e–f), and these observed features are also well captured by the model.

On 18 March, the P-3B conducted observations mainly over the East China Sea and the Yellow Sea. This flight between Okinawa and Yokota, an airbase near Tokyo (*cf.* Fig. 2c), was planned to include a northbound leg along 124.5°E beginning at 30°N stretching into the Yellow Sea up to 37°N, then south again before heading east toward Yokota. Figures 5a–d show that specific humidity was low and north-westerly winds prevailed in the lower troposphere. Under such conditions very high concentrations of NO, NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> as shown in Figs. 5e–h were observed at ~1210 and 1400 CST over the East China Sea and the Yellow Sea. Concentrations of ANO<sub>3</sub> and ANH<sub>4</sub> were much higher comparing with those obtained during the flights conducted on 13 and 17 March (*cf.* Figs. 3g–h and Figs. 4g–h). From Fig. 5 we can see that the model results are generally in good agreement with the observations.

The flight of the P-3B to the east of Japan was conducted on 21 March. It started observations at about 0830 CST, and firstly headed south-eastward and then eastward, turned north and flew along 150°E, finally returned to Yokota (*cf.* Fig. 2d). Figures 6f–h show that high levels of NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> were observed at ~1430 CST and in the region close to Tokyo, and the model reproduced these high values quite well while ANO<sub>3</sub> was greatly underestimated at ~1430 CST. From Fig. 6 we find that modelled meteorological parameters and chemical species are generally in good agreement with their observed ones except for wind directions and ANO<sub>3</sub> at ~1430 CST. The observed winds were

Table 1. Statistical summaries of model results and the observations along the P-3B flights 12–15.

	QV <sup>c</sup>	TA <sup>b</sup>	WS <sup>a</sup>	WD <sup>a</sup>	NO <sup>b</sup>	NO <sub>2</sub> <sup>b</sup>	ANO <sub>3</sub> <sup>b</sup>	ANH <sub>4</sub> <sup>b</sup>
N <sup>c</sup>	349	349	349	349	316	244	190	280
CO <sup>d</sup>	3.52	275.3	13.35	247.5	35.7	117.0	2.00	1.05
CM <sup>d</sup>	3.99	274.8	13.72	240.9	30.5	122.3	1.63	1.79
σ <sub>O</sub> <sup>d</sup>	3.63	10.5	8.57	75.7	54.0	214.6	5.38	2.07
σ <sub>M</sub> <sup>d</sup>	3.57	10.1	8.06	71.0	21.8	135.7	3.32	1.88
R <sup>e</sup>	0.96	0.99	0.94	0.66	0.48	0.44	0.82	0.83

<sup>a</sup>QV, TA, WS, and WD represent for wind speed (m/s), wind direction (°), temperature (°K) and specific humidity.

<sup>b</sup>Units of NO, NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> are pptv, pptv, μg/m<sup>-3</sup> and μg/m<sup>-3</sup>, respectively.

<sup>c</sup>N is the number of paired samples.

<sup>d</sup>CO, σ<sub>O</sub>, CM and σ<sub>M</sub> are the average value and the standard deviation of observed and modeled quantities.

<sup>e</sup>R stands for the correlation coefficient between observed and modeled quantities.

north-easterly while the simulated were north-westerly, which might lead to the great underestimation of ANO<sub>3</sub>.

The above analysis and the statistics summaries presented in Table 1 show that meteorological parameters exhibited strong variations in time and space over East Asia, for example, horizontal wind speed and direction changed significantly in the lower atmosphere, while westerly winds prevailed in the layers above ~4 km and winds were much stronger at mid-latitudes than at low latitudes (*cf.* Figs. 3c–d and 4c–d), and as shown in Table 1, the standard deviation (SD) of specific humidity (3.63 g/kg) is comparable to its average value (3.53 g/kg), and the standard deviations of wind speed and wind direction are as large as 8.57 m/s and 75.73°. From Fig. 3–6 and Table 1 we can see that the model reproduced the time spatial variations in specific humidity, temperature, wind speed and direction quite well, the average values of the modelled parameters are all in good agreement with their observed ones. The correlation coefficients between model results and observations are 0.96, 0.99, 0.94 and 0.66, respectively. Although the correlation coefficient for wind directions is relatively low, the percentage of the comparison points at which the absolute bias is within 30° is 79.9% if we use a value of 30° as the desired accuracy criteria for wind directions (Cox et al., 1998). As the observations used in this study were obtained over the western Pacific, further evaluation of the model's ability in producing local circulation and boundary structure over urban areas and mountainous regions needs to be done.

Fig. 3–6 also show that the concentrations of NO, NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> have strong tempo-spatial variations with high values in the lower troposphere. That is because NO<sub>x</sub> and NH<sub>3</sub> are mostly emitted in the continental boundary layer, and NO, NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> levels over the water areas strongly depend on the transport and transformation processes. Comparing the

observed concentrations in the layers below  $\sim 2$  km on different days and over different areas, we find that  $\text{NO}_2$  concentrations were generally much higher over the Yellow Sea and the East China Sea on 17 and 18 March (*cf.* Figs. 3f and 4f) than over the South China Sea on 13 March (*cf.* Fig. 2f), and the mixing ratios of  $\text{ANO}_3$  and  $\text{ANH}_4$  on 18 March (*cf.* Figs. 4g–h) greatly exceeded those on 17 March (*cf.* Figs. 3g–h) over the Yellow Sea and the East China with nearly the same  $\text{NO}$  and  $\text{NO}_2$  levels. From these figures we also find that the model captured these observed features well. Table 1 indicates that average simulated values of  $\text{NO}$ ,  $\text{NO}_2$ ,  $\text{ANO}_3$  and  $\text{ANH}_4$  generally agree with their observed ones, and correlation coefficients for  $\text{ANO}_3$  and  $\text{ANH}_4$  are larger than 0.80, while simulated standard deviations are much smaller than observed ones and correlation coefficients for  $\text{NO}$  and  $\text{NO}_2$  are 0.48 and 0.44. The model's inability to reproduce very high levels of  $\text{NO}$  and  $\text{NO}_2$  maybe due to the coarse resolution of the model system and missing  $\text{NO}_x$  emissions, and this inability may lead to the underestimation of maximum  $\text{ANO}_3$  and  $\text{ANH}_4$  concentrations and low correlation coefficients for  $\text{NO}$  and  $\text{NO}_2$ .

### 3.2. Geographical and seasonal distributions of $\text{ANO}_3$ concentrations

Figure 7 shows the geographical distributions of  $\text{ANO}_3$  mixing ratios in East Asia in January, April, July and October in

the boundary layer (approximately 0–1000 m) as calculated by CMAQ. Results are presented as monthly averages. The  $\text{ANO}_3$  concentrations have noticeable differences between the four months. Generally the concentrations are higher in the winter, spring and fall, and lower in the summer and seasonal variations are typically strongest over China. As nitrate aerosols in the model normally present as particulate ammonium nitrate ( $\text{NH}_4\text{NO}_3$ ), formed from the reversible reaction of gas phase ammonia ( $\text{NH}_3$ ) and nitric acid ( $\text{HNO}_3$ ), and  $\text{HNO}_3$  is easily absorbed by rain water and cloud water, so the seasonal variation in  $\text{ANO}_3$  concentrations is heavily influenced by the seasonal changes in precipitation and temperature.

Figure 7a shows that in January elevated  $\text{ANO}_3$  concentrations ( $\geq 2 \mu\text{g}/\text{m}^3$ ) are mainly found over eastern China and its coastal areas with maximum concentrations in an area over Central China between the Yangtze River and the Yellow River. This area is characterized by high emissions of  $\text{NO}_x$  and  $\text{NH}_3$  (Streets et al., 2003; Benkovitz et al., 1996). In the wintertime the photochemical activity is relatively weak, but  $\text{HNO}_3$  is formed through heterogeneous reactions on the surface of aerosols (Schaap et al., 2003). January is under the influence of the winter monsoon in East Asia due to strong pressure gradients between the Siberian (Continental) High and Okhotsk (Maritime) Low, and the precipitation is very low over China but is substantial over Japan. During the peak of the monsoon, there is strong subsidence in the major part of the East Asia continent (Ding, 1994), and

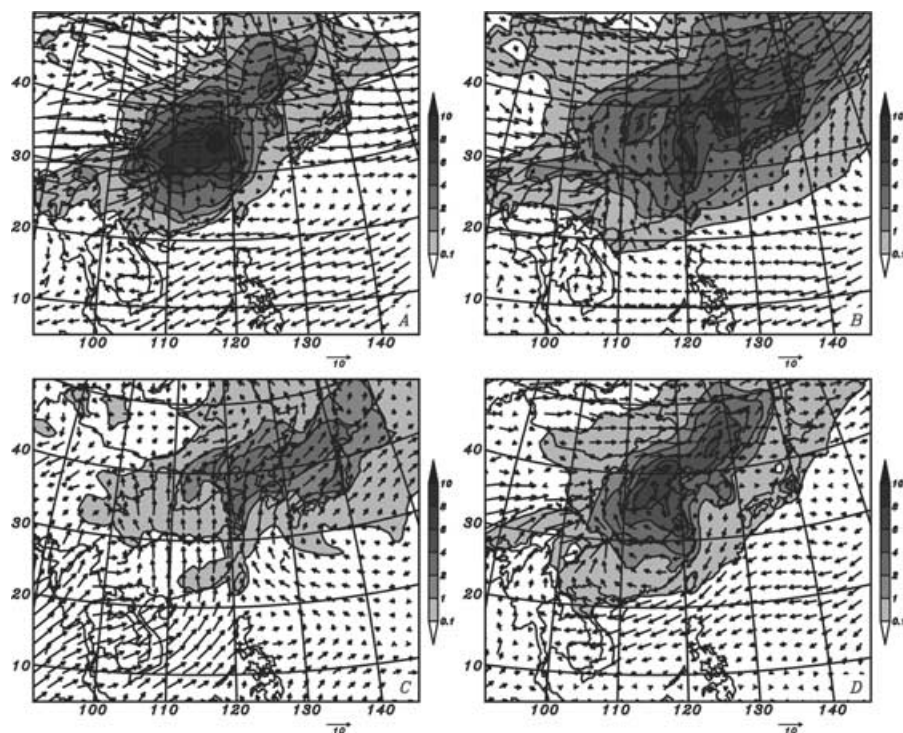


Fig. 7. Horizontal distributions of monthly averaged  $\text{ANO}_3$  concentrations ( $\mu\text{g}/\text{m}^3$ ) in the boundary layer (from surface to  $\sim 1000$  m) in (a) January, (b) April, (c) July and (d) October of 2001. Also shown are monthly averaged wind vectors at an altitude of  $\sim 500$  m.

pollution is trapped in the boundary layer. The monthly averaged wind flow pattern shown in Fig. 7a presents the typical pollutant transport pathway in association with the continental anticyclones and cold air outbreaks.

Spring in East Asia corresponds to the period of transition between the winter and summer monsoons (Ding, 1994), and strong north-westerly continental outflow occurs in April in the mid and upper troposphere. By March the Siberian High weakens, while the Pacific high-pressure system is building up. As a result, the strength of the winter monsoon winds and the frequency of occurrence of the cold south moving surges decrease, while incursions of warmer and tropical air from the south become frequent. As shown in Fig. 7b, the flow in the boundary layer is onshore over China south of about 30°N but strong westerlies prevail north of 40°N. The convergence zone over central and eastern China, where air masses from the north driven by monsoon winds encounter oceanic air masses from the south, plays an important role in the springtime export of pollution from the Asian continent (Zhang et al., 2003). Figure 7b shows a wide area with ANO<sub>3</sub> concentrations larger than 2  $\mu\text{g}/\text{m}^{-3}$  covering eastern China, Korea peninsula and Japan and part of the Yellow Sea and the Sea of Japan, and the transport processes played an important role in leading high concentrations over the region including the Yellow Sea and the Sea of Japan, and the water areas to the north-east of Japan. High levels of ANO<sub>3</sub> ( $\geq 2 \mu\text{g}/\text{m}^{-3}$ ) also attribute to stronger photochemical activity than in January.

July is in the middle of the summer monsoon season, westerly winds weaken and southerly monsoons dominate as far north as 30°N. The meteorological situation is characterized by predominant low level southerly and south-westerly circulation and a maximum in precipitation and cloudiness. The southerly winds push pollution northwards and north-eastwards. In the summer strong solar radiation leads to much more convection in the surface layer and increase of dry deposition of ANO<sub>3</sub>. Model results show that the distribution pattern of dry deposition is quite similar to that of monthly averaged airborne concentration in the boundary layer presented in Fig. 7, and deposition amount increases from January to July. Figure 7c shows that the maximum concentrations are below  $\sim 4 \mu\text{g}/\text{m}^{-3}$  due to the higher cloud water contents, precipitation and temperatures and stronger dry deposition in July than in January.

October is in the early stage of the dry season over East Asia, and strong north-westerly continental outflow occurs in the mid and upper troposphere. Figure 7d shows anticyclone circulations over central and eastern China and prevailing westerlies and south-westerlies north of 40°N. The anticyclone circulations imply that much of the pollution originating in central and eastern China will flow clockwise out to sea then return to the south-west. While the westerly and south-westerly winds north of 40°N sweep pollution north-eastward. From Fig. 7d we can see that high ANO<sub>3</sub> concentrations ( $\geq 2 \mu\text{g}/\text{m}^{-3}$ ) are found over eastern China and its coastal areas north of 30°N. The general

distribution pattern shown in Fig. 7d is similar to that in January, but maximum concentrations are smaller (*cf.* Fig. 7a) because of the subsidence over the continent is stronger in January in association with the winter monsoon.

#### 4. Summary

The geographical and seasonal characteristics of nitrate aerosol concentration distributions in the boundary layer over East Asia in the year of 2001 were investigated by use of CMAQ with meteorological fields from RAMS. For evaluating the model performances, the modelled values of wind direction, wind speed, temperature and specific humidity and concentrations of NO, NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> were compared with the observations obtained onboard the P-3B in the period 13 to 21 March 2001, when the P-3B performed four consecutive flights over four different regions: the South China Sea, the East China Sea, the Yellow Sea and the water areas to the east of Japan. Comparison shows that the observed wind direction, wind speed, temperature and specific humidity exhibited strong time spatial variations and the model captured these features. The modelled values are generally in good agreement with the observed ones, and the correlation coefficients are 0.66, 0.94, 0.99 and 0.96, respectively. If a value of 30° is used as the desired accuracy criteria for wind directions, the percentage of the comparison points at which the absolute bias is within the criteria reaches 79.9%. As the observations used in this study were obtained over the western Pacific, further evaluation of the model's ability in producing local circulation and boundary structure over urban areas and mountainous regions needs to be done.

The observations indicate that the mixing ratios of NO, NO<sub>2</sub>, ANO<sub>3</sub> and ANH<sub>4</sub> strongly depended on their transport and transformation processes, and large concentrations were mostly detected in the layers below  $\sim 2$  km. The comparison shows that the model generally reproduced the concentration variation patterns along the flight tracks. The simulated average values agreed reasonably well with their observed ones, and correlation coefficients for ANO<sub>3</sub> and ANH<sub>4</sub> are larger than 0.80, while simulated standard deviations are much smaller than observed ones and correlation coefficients for NO and NO<sub>2</sub> are 0.48 and 0.44. Analysis indicates that the model was unable to reproduce very high levels of NO and NO<sub>2</sub> due to the coarse resolution of the model system and missing NO<sub>x</sub> emissions, and this inability may lead to the underestimation of maximum ANO<sub>3</sub> and ANH<sub>4</sub> concentrations and low correlation coefficients for NO and NO<sub>2</sub>.

Analysis of horizontal distributions of monthly averaged ANO<sub>3</sub> concentrations in the boundary layer in January, April, July and October of 2001 shows that the ANO<sub>3</sub> mixing ratios have noticeable differences between the four months. Generally the concentrations are higher in the winter, the spring and the fall and low in the summer, and seasonal variations are typically strongest over China. Beside high emissions in NO<sub>x</sub> and NH<sub>3</sub>, the maximum concentrations over central and eastern China in



January are due to the strong subsidence associated with the winter monsoon. Long-range transport may cause elevated concentrations in remote areas under favourable meteorological conditions, but the seasonal variation in the  $\text{ANO}_3$  concentration is heavily influenced by the seasonal changes in precipitation and temperature.

## 5. Acknowledgments

This study was supported by grants from the Ministry of Science and Technology of China (Grant No. 2007CB407303 and 2005CB422205) and the National Natural Science Foundation of China (Grant No. 40333029). The authors gratefully acknowledge the team members of TRACE-P for the observational data.

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